

DYNAMIC PCDD/F LEVEL MODELING USING STELLA SIMULATION IN TAIWAN

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1. Introduction

Food intake takes over 95% of PCDD/Fs exposure for human. Therefore, the PCDD/Fs levels in food and the human intake rate are crucial factors to human PCDD/Fs exposure. Due to its high lipophilicity (i.e., octanol/water partition coefficient) and low water solubility, once PCDD/Fs deposit from atmosphere to land surface and ingested by biota, they tend to adsorb strongly to animal's lipid tissue and accumulate along the food chain, which is called bioaccumulation. PCDD/Fs emissions can transport from one country to another, therefore, high levels are found even in remote polar area. In order to estimate PCDD/Fs levels in biota, this study focuses on the intermediate movement and the mass balance of PCDD/Fs among environmental media, rather than addressing levels of different PCDD/Fs congeners or their transformation mechanisms, such as oxidation, photolysis, or biodegradation. The model's domain included the terrestrial territory of 35,980 km², including area of 500 km² under freshwater bodies, and also Taiwan's exclusive economic zone of 180,000 km² (Figure 1). This domain was divided into four compartments—air, soil, freshwater, and seawater—PCDD/Fs levels in both terrestrial and aquatic biota could be estimated by the bioaccumulation factors calculated in this study.



Figure 1 Model domain including exclusive economic zone for Taiwan.

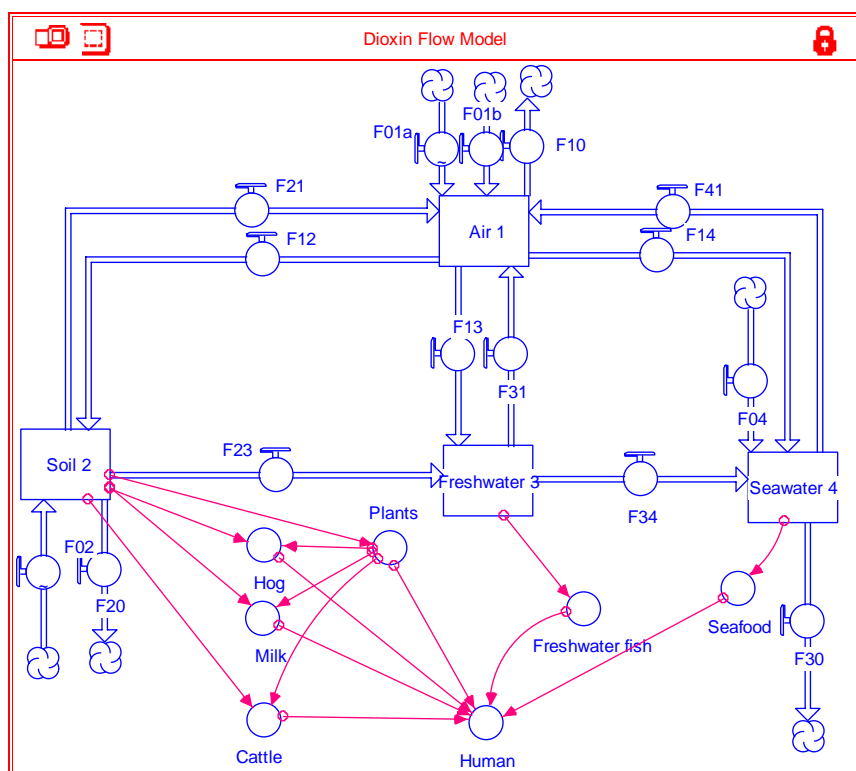
International toxic equivalent (I-TEQ), calculated by toxic equivalency factors (TEFs), is a unit for health risk assessment¹. TEF value for the most toxic dioxin, 2,3,7,8-TCDD, is assigned as 1 and 0–0.5 for other PCDD/Fs congeners and dioxin-like PCBs. This study used I-TEQ as units for PCDD/Fs levels measured in the environmental media and biota for the consequent human exposure assessment.

2. Materials and methods

Model Construction. Figure 2 demonstrates the Dynamic Dioxin Flow model, which was depicted using the STELLA v.9.0 (iSee™ systems, USA) and was modified from a model for PCBs in a river basin².

Initial Reservoir Mass. The PCDD/Fs mass in each environmental reservoir in the beginning of simulation year 1979 was estimated by multiplying the concentration with the volume, which is derived as a product of the area and the depth of the environmental reservoir. The depths of atmosphere, soil, freshwater, and seawater were assumed to be 5,000 m, 0.2 m, 5 m, and 50 m, respectively. Due to the fact that there are no domestic measurements for PCDD/Fs in 1979, present background levels from Taiwan EPA were used as surrogate.

According to the data from Taiwan EPA, observed PCDD/Fs background concentrations in air, soil, freshwater and seawater approximated 0.341 pg I-TEQ/Nm³, 3.89 pg I-TEQ/g, 0.043 pg I-TEQ/L, and 0.019 pg I-TEQ/L, respectively.



Note: The four PCDD/Fs reservoirs—air, soil, freshwater, and seawater—were numbered 1, 2, 3 and 4, and the input/output fluxes between them were represented by F_{ij} , where $i, j = 1, 2, 3, \text{ or } 4, i \neq j$.

Figure 2 The Dynamic Dioxin Flow model.

Anthropogenic Emissions. Trends in the anthropogenic PCDD/Fs emissions have been reported by Taiwan EPA since 1999. Emissions before 1999 were estimated in this study based on emission factors and activities for electric arc furnaces, steel smelters, and municipal waste incinerators (MWIs), which were major (70%) contributors, from 1979 to 1999. Figure 3 shows that anthropogenic emissions increased from 34.04 g I-TEQ/yr in 1979, climaxed at 327.45 g I-TEQ/yr in 2002, and then decreased to 52.8 g I-TEQ/yr in 2009. Emissions after 2009 were predicted to grow slightly for each year till 2029 (Figure 3).

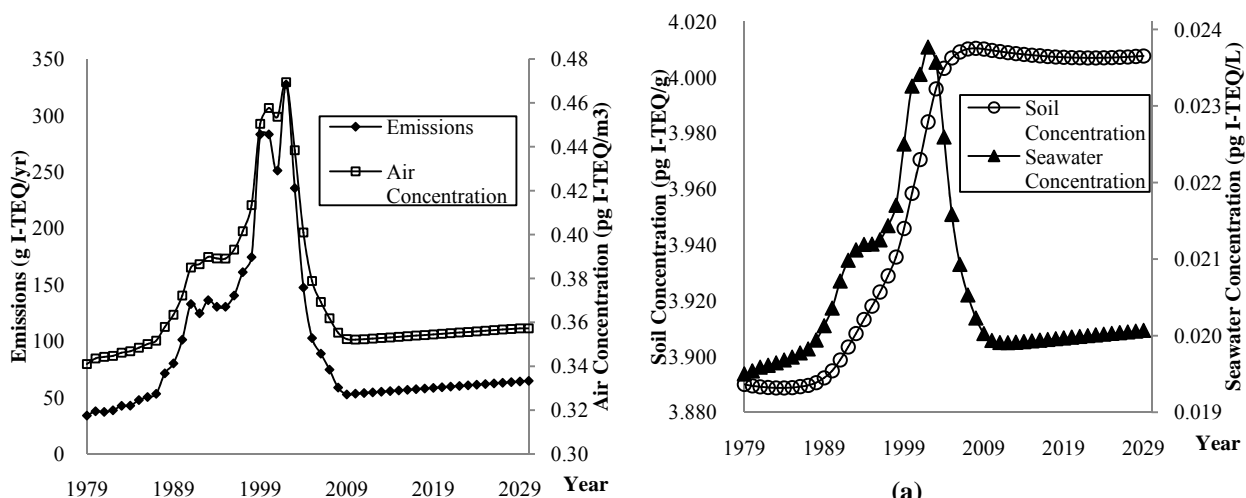
Natural Emissions. Natural PCDD/Fs emissions are correlated with re-emissions from land and water surface. Once deposited to soil, PCDD/Fs bound strongly to colloids and particles and volatilization half-life from soil may be as long as 12 years. Most of PCDD/Fs occurring in water are expected to be associated with sediments or suspended materials. Due to its adsorption to sediment, the removal half-life may be over 50 years. The volatilization rates were determined as 0.027–0.47 pg I-TEQ/m²-hr, depending on air-surface concentration gradients³. The evaporation duration approximated 225 days yr⁻¹ based on meteorological data.

Wet and Dry Deposition. PCDD/Fs emitted to the atmosphere can be removed by wet and dry deposition. Wet deposition flux of PCDD/Fs is associated with both vapor dissolution into rain and the removal of suspended particulates by precipitation. PCDD/Fs levels in raindrops can be estimated by atmospheric PCDD/Fs concentration and both gas and particle scavenging ratios, which were determined as 741 and 4.2E+4 (dimensionless)⁴. Average atmospheric PCDD/Fs concentrations approximated 6.2 fg I-TEQ/m³ on the marine and

51 fg I-TEQ/m³ on the land^{5,6}. These concentrations were assigned for gas and particle phase using a constant gas/particle partitioning factor of 0.25⁷. The estimated wet deposition rates were 200 pg I-TEQ/m²-yr on the marine and 4,303 pg I-TEQ/m²-yr on the land. Dry deposition involves dry particulate deposition by gravitational settling or turbulent diffusion and vapor phase deposition by diffusion. The deposition flux is derived as the product of the vapor/particle phase concentrations and the deposition velocity of vapor or particle, which were 0.01 cm/s or 0.3–0.44 cm/s, respectively^{8,9}. The estimated dry deposition rates were 242 pg I-TEQ/m²-yr on the marine and 2,805 pg I-TEQ/m²-yr on the land. The total deposition rates were estimated as 0.44 ng I-TEQ/m²-yr on the marine and 7.11 ng I-TEQ/m²-yr on the land.

Dynamic Modeling. A static mass balance, that is, the sum of PCDD/Fs inflow fluxes equal to the outflow, was set for each of the four environmental reservoir—air, soil, freshwater and seawater—in the initial year 1979. The parameterization for dynamic modeling in this study was simply assumed that there were linear relationships between atmospheric deposition rate and reservoir mass and between reemission rate and reservoir level. Consequently, the variables of dimensionless coefficients were derived. The simulation used a *dt* of 1 day, and a duration of 50 years (1979–2029).

Bioaccumulation. PCDD/Fs enter the terrestrial food chain by deposition onto plant surfaces or the soil which are subsequently adsorbed by plants via roots or ingested by animals. The variables of uptake rate and metabolism factors for plant and animals were drawn from the US EPA algorithms¹⁰. As shown in Figure 2, the PCDD/Fs levels for agricultural crops and livestock, such as hog, cattle, and milk, commonly ingested by residents in Taiwan, are associated with both soil and plants concentrations. Fish can strongly bioaccumulate PCDD/Fs through aquatic food chain. The biomagnification factor (BMF) is the ratio of PCDD/Fs level in fish tissue to its concentration in the water, and the BMF values for the freshwater and seawater fish were set as 5,800 and 58,660, respectively, suggested by US EPA¹¹.



(b) Figure 3 Temporal trends of the (a) annual anthropogenic emissions and the simulated annual average PCDD/Fs concentration for air, (b) freshwater, and seawater.

3. Results and discussion

Results. As shown in Figure 3, the PCDD/Fs levels for air and seawater are consistent with the trends of anthropogenic emission. However, levels for surface soil did not significantly decline while the emission dramatically decreased during 2002 and 2009, attributing to regulations for electric arc furnaces and MWIs emission standards. It is mainly because the soil keeps the bulk of PCDD/Fs among the reservoirs¹², so its concentration tends to be relatively stable. By comparing the simulated PCDD/Fs concentrations with the observed (Table 1), it shows that there was acceptable corroboration between simulated and observed values. This

result reveals that the Dynamic Dioxin Flow model is pretty accurate on predicting PCDD/Fs concentrations not only for environmental media, but also for aquatic and terrestrial food chain.

Table 1 Comparison of simulated and observed PCDD/Fs concentrations in various media.

Media (units)	Time	Observed	Simulated
air(pg I-TEQ/m ³)	2008	0.21 ^a	0.36
soil(pg I-TEQ/g)	2005	3.89 (0.254–15.2) ^a	3.95
freshwater(pg I-TEQ/L)	2007	0.044 (0.008–0.079) ^a	0.045
seawater(pg I-TEQ/L)	2008	0.019 ^a	0.020
plants(pg I-TEQ/g)	2003	1.33(for pasture sampled) ^b	1.81
beef(pg I-TEQ/g)	2004	0.967±0.692 ^b	0.973
milk(pg I-TEQ/g)	2004	1.34±0.84 ^b	0.256
	2010	0.564 ^b	0.256
pork (pg I-TEQ/g)	2004	0.306±0.214 ^b	0.128
freshwater fish(pg I-TEQ/g)	2004	1.850±0.36 ^b	0.319
seawater fish(pg I-TEQ/g)	2004	4.700±3.82 ^b	1.297

^aTaiwan Environmental Protection Administration. ^bTaiwan Ministry of Health.

Discussion. The Dynamic Dioxin Flow model developed by the STELLA software is low-cost, simple and fast. Moreover, this model showed good performance on current concentrations for both environmental media and biota. On health risk or exposure assessment, it could be a substitute for complex environment model, such as atmospheric dispersion model for anthropogenic emission and water quality model for rivers, especially in long-term prediction, such as modeling fish PCDD/Fs level after dozens of years or even one hundred years later. However, parameterization for dynamic modeling in this study should be more sophisticated, such as, in the calculation of dynamic atmospheric deposition rate and the consideration of sediment burial, degradation and photolysis. This model (Figure 2) illustrates the human PCDD/Fs exposure for a more comprehensive study related to health risk assessment, which could be carried out by incorporating parameters like food intake rates for residents in different countries.

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