

STRUCTURE ELUCIDATION OF 2-*endo*,3-*exo*,5-*endo*,6-*exo*,8,8,10-HEPTACHLOROBORNANE, AN ABUNDANT TOXAPHENE COMPONENT IN TOP PREDATORS AND SEDIMENTS

Walter Vetter¹, Elke Scholz¹, Bernd Luckas¹, Keith A. Maruya²

¹ Friedrich-Schiller-University Jena, Department of Food Chemistry, Dornburger Str. 25, D-07743 Jena, Germany

² Skidaway Institute of Oceanography, 10 Ocean Science Circle, Savannah, GA, 31411, USA

Introduction

Toxaphene was the most heavily used organochlorine pesticides in the U. S.¹. The global production was estimated at 1.3 million tons². Despite its ban in many countries over a decade ago, elevated concentrations of compounds of technical toxaphene (CTTs) are still found in the environment. Presently, CTTs are a major organochlorine contaminant of aquatic life worldwide³. However, only a subset of the several hundred CTTs in the original mixture are detectable in environmental samples⁴. In a previous study on Antarctic seals, the GC/ECNI-MS-SIM abundance of a heptachloro component designated as "7-1" was >10% of B8-1413 (P-26), a known, persistent CTT⁵. The purpose of this study was to elucidate the structure of 7-1.

Materials and Methods

Sample origin and reference standards. 7-1 was isolated from toxaphene-contaminated estuarine sediment⁶. A solution of the technical product Camphechlor (10 ng/ μ L) was obtained from Promochem (Wesel, Germany). Standard solutions of CTTs were obtained from Dr. Ehrenstorfer (Augsburg, Germany), Promochem, or isolated in our lab^{7,8}. Both systematic AV-codes⁹ and Parlar numbers¹⁰ are used to name individual CTTs.

Isolation of 7-1 from sediment. CTTs including 7-1 were extracted from 1 kg sediment by shaking with hexane/acetone (1:1, v:v) overnight. The organic layer was partitioned into n-hexane and then treated with H₂SO₄ and acid-activated copper. The sediment extract (~1 mg toxaphene) was fractionated on 60 g silica, eluting with n-hexane. Fraction 325-350 mL containing the bulk of 7-1 was evaporated to dryness and re-dissolved in 500 μ L acetonitrile. This extract was injected into a RP-HPLC system⁷. The HPLC fraction eluting between 12.5-13.0 min contained >90% 7-1.

Sample clean-up. Matrix interferences in biological samples were removed from the organochlorine fraction by digestion with acids, liquid-liquid extraction with n-hexane, repeated treatment with sulfuric acid, and adsorption chromatography on silica¹¹.

Gas chromatography/electron capture negative-ionization mass spectrometry (GC/ECNI-MS). Measurements were performed with a HP 5890 Series II plus/5989B GC/MS system using previously published parameters¹². A 63 m x 0.25 mm i. d. column coated with 0.25 μ m of CP-Sil 2 (Chrompack, The Netherlands) or a 30 m x 0.25 mm i. d. column coated with 0.2 μ m of 25% randomly *tert.*-butyldimethylsilylated β -cyclodextrin diluted in PS086 (BGB Analytik, Switzerland) was used. In the SIM mode we monitored *m/z* 343, 345, 378, and 380 for 7-1.

¹H-NMR measurements. The ¹H-NMR measurements of 7-1 were performed on a Bruker DRX

500 spectrometer. All chemical shifts were referenced to the solvent peak (CDCl_3) and recalculated with respect to TMS ($\delta(^1\text{H}) = 7.260$ ppm for CDCl_3). Two dimensional shift correlated and phase-sensitive, time proportional phase incrementation spectra (COSY and NOESY, respectively) were recorded with a mixing time of 220 ms. Acquisition times for COSY and NOESY were 20.5 and 70.5 h, respectively. $^1\text{H-NMR}$ -data for 7-1 were as follows: H8 (CHCl_2), 6.68 ppm, s; H2-*exo*, 5.12 ppm, dd, 5.5 Hz and ~ 1.2 Hz; H3-*endo*, 4.66 ppm, m, 5.5 Hz; H6-*endo*, 4.56 ppm, ? Hz, ? Hz; H5-*exo*, 4.56 ppm, m, (to 2.68 ppm); H10a (CH_2Cl), 4.19 ppm, dd, 12.8 Hz, ? Hz; H10b (CH_2Cl), 3.57 ppm, d, 12.8 Hz; H4, 2.68 ppm, m, ~ 1.2 Hz (to 4.56); H9 (CH_3), 1.89 ppm, s.

Results and Discussion

Isolation and NMR structure elucidation. Based on ECD response factors for several early eluting CTTs, we estimated the isolated mass of 7-1 to be ~ 20 μg . Although chemical shifts of two protons were identical and some coupling constants could not be resolved (see above), the structure of 7-1 was unequivocally determined to be 2-*endo*,3-*exo*,5-*endo*,6-*exo*,8,8,10-heptachlorobornane (B7-1000) based on a comparison of the $^1\text{H-NMR}$ data with that for B8-1413 (P-26). Except for the additional proton H10 on B7-1000 (the signal of B8-1413 (P-26) at 6.42 ppm is distributed in 3.57 and 4.19 ppm), Δppm of all other protons matched that of B8-1413 (deviations of 0.01 to 0.17 ppm). B8-1412 also showed similar chemical shifts as B7-1000 for the two protons on C10⁷. Further confirmation of the structural assignment was derived from molecular modeling as reported elsewhere¹³.

GC/ECNI-MS. The full scan mass spectrum of B7-1000 was dominated by the molecular ion starting at m/z 376 (Figure 1). Formation of the molecular ion is unusual for heptachlorobornanes, which usually show highest abundance for the $[\text{M-Cl}]$ fragment ion¹⁴. The $[\text{M-Cl}]$ fragment ion at m/z 341 was overlapped by the $[\text{M-HCl}]$ fragment ion which accounted for $\sim 40\%$ of the former (i.e. $[\text{M-Cl}]$) fragment ion. A large relative contribution of the $[\text{M-HCl}]$ fragment ion was also found for B8-1413 (P-26), providing further evidence as to the similarity of the two CTTs.

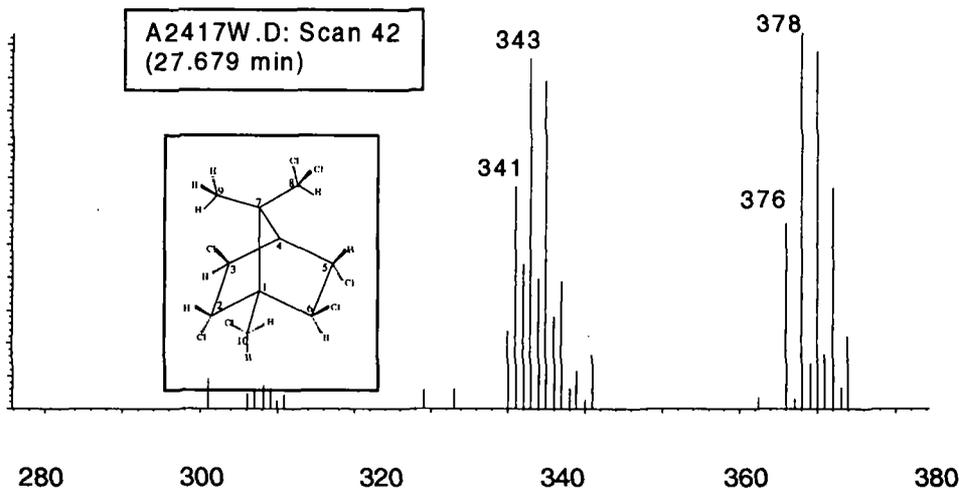


Figure 1: GC/ECNI mass spectrum and structure of B7-1000 (7-1)

TOXAPHENE - POSTERS

Significance of B7-1000 in environmental samples. GC/ECNI-MS in the SIM mode was used to identify B7-1000 in technical toxaphene and environmental samples (Figure 2). B7-1000 is the first eluting heptachlorobornane in toxaphene on non-polar GC phases (e.g. DB-5). GC/MS analysis of technical toxaphene is therefore a suitable method for the identification of B7-1000. Because a quantitative solution was not available and ECNI-MS response factors can vary significantly for chlorobornanes, we cannot provide concentrations for B7-1000. Preliminary results suggest, however, that the detector response ratios of B7-1000 and B8-1413 (P-26) using ECD and ECNI-MS are similar. Under this assumption, B7-1000 accounted for ~10 % of the ECNI abundance of B8-1413 (P-26) in most of our samples (Table 1).

Table 1: ECNI-MS signal intensity of B7-1000 relative to B8-1413 (P-26)

Origin, Species	Sample size	% of B8-1413 (P-26) mean (range)
Iceland, harbor porpoise (<i>Phocoena phocoena</i>)	n = 4	9.2 (8.3 – 10.7)
Spitsbergen, harp seal (<i>Phoca sibirica</i>)	n = 9	10.0 (pool sample)
Antarctic, Weddell seals (<i>Leptonychotes Weddelli</i>)	n = 8	16.2 (11.6-25.6)

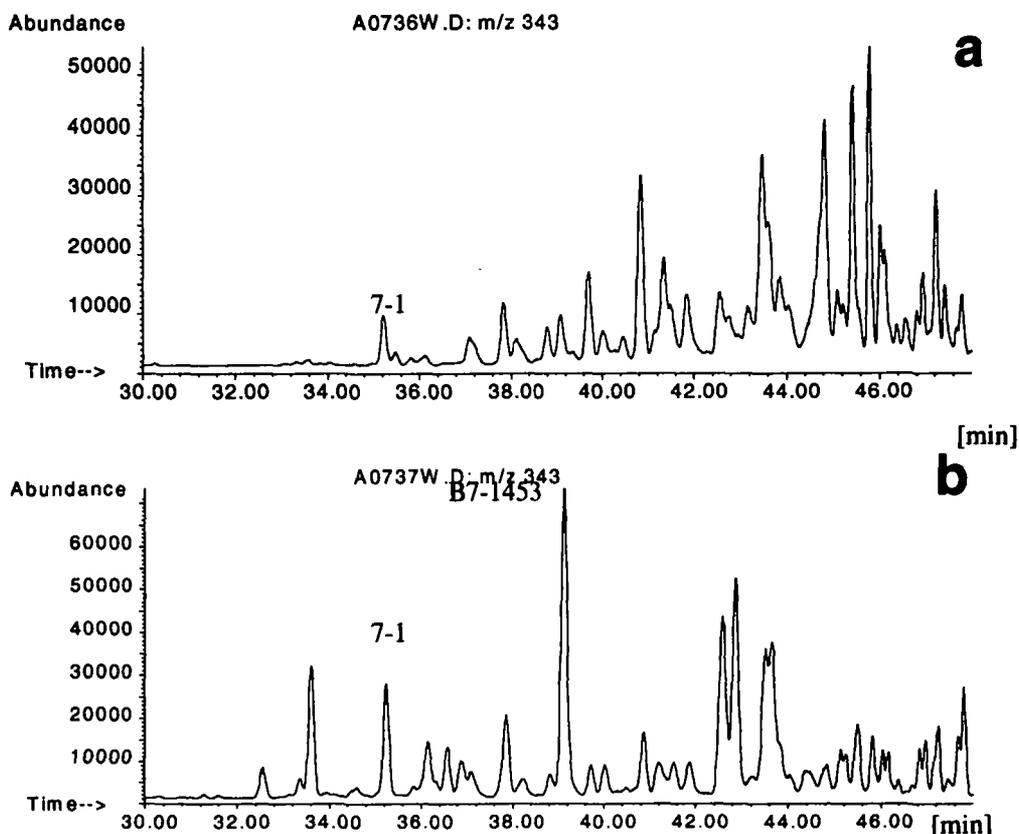


Figure 2: GC/ECNI-MS detection of B7-1000 in (a) technical toxaphene and (b) blubber of harbor porpoise (*Phocoena phocoena*). Unlabelled peaks in (b) are not chlorobornanes.

ORGANOHALOGEN COMPOUNDS

TOXAPHENE - POSTERS

We also detected B7-1000 in human milk, adipose tissue of a polar bear, and fish samples (salmon fillet and cod liver). It is thus surprising that B7-1000 has not been reported more frequently in the literature. Monitoring a time window for m/z 343 that is too narrow is one possibility. Using non-selective methods such as GC/ECD, a pre-separation of PCBs should be performed before quantification of CTTs. Many authors have reported that some CTTs (particularly B8-1413 (P-26)) might smear into the PCB fraction^{15,16}. Note that B7-1000 elutes earlier from silica than B8-1413 (P-26), and thus will elute in the PCB fraction generated using standard chromatographic clean-up protocols.

Conclusion. The structure elucidation of previously unidentified chlorobornanes like B7-1000 enhances our understanding of the environmental fate of toxaphene. Many environmentally relevant CTTs are currently available as reference standards, enabling the analyst to quantify CTTs on a congener-specific basis. It should be emphasized, however, that none of the commercially available heptachlorobornanes are regularly detected in higher organisms. Thus, the commercial availability of B7-1000 would be an important contribution to the analysis of toxaphene.

Acknowledgment. The authors thank U. Gräfe and H. Heineke (HKI, Jena, Germany) for NMR-experiments, T. Walters and K. Smalling of Skidaway Institute of Oceanography, and W. Quinn of Hercules Inc. for sampling and extraction of the sediment.

References

1. Casida J.E., Holmstead R.L., Khalifa S., Knox J.R., Ohsawa T., Palmer K.J. and Wong R.Y. (1974) *Science* 183, 520.
2. Voldner E.C. and Li Y.-F. (1995) *Sci. Total Environ.* 160/161, 201.
3. Bidleman T.F., Walla M.D., Muir D.C.G. and Stern G.A. (1993) *Environ. Toxicol. Chem.* 12, 701.
4. Vetter W. and Oehme M.; Toxaphene. Analysis and environmental fate of congeners. *The Handbook of Environmental Chemistry*, Vol. 3, Part K: New Types of Persistent Halogenated Compounds (Paasivirta J., Ed.). Springer, Berlin Heidelberg, 1999, pp. 237-287.
5. Vetter W., Krock B. and Luckas B. (1997) *Chromatographia* 44, 65.
6. Vetter W. and Maruya K. (2000) *Environ. Sci. Technol.* 34, 1627.
7. Vetter W., Klobes U., Krock B., Luckas B., Glotz D. and Scherer G. (1997) *Environ. Sci. Technol.* 31, 3023.
8. Krock B., Vetter W., Luckas B. and Scherer G. (1996) *Chemosphere*, 33, 1005.
9. Andrews P. and Vetter W. (1995) *Chemosphere* 31, 3879.
10. Parlar H., Angerhöfer D., Coelhan M. and Kimmel L. (1995) *Organohalogen Compd.* 26, 357.
11. Luckas B., Vetter W., Fischer P., Heidemann G. and Plötz J. (1990) *Chemosphere* 21, 13.
12. Vetter W., Klobes U., Krock B. and Luckas B. (1997) *J. Microcol. Sep.* 9, 29.
13. Vetter W. and Scherer G. (1999) *Environ. Sci. Technol.* 33, 3458.
14. Swackhamer D.L., Charles M.J. and Hites R.A. (1987) *Anal. Chem.* 59, 913.
15. Alder L. and Vieth B. (1996) *Fresenius J. Anal. Chem.* 354, 81.
16. Muir D.C.G., Ford C.A., Grift N.P., Metner D.A. and Lockhart W.L. (1990) *Arch. Environ. Cont. Toxicol.* 19, 530.