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Levels of Toxic Polychlorinated Biphenyls , Dibenzo-p-dioxins and Dibenzofurans in Biotic and Abiotic Samples from Galveston Bay, Texas U.S.A.

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Introduction

A vast number of anthropogenic compounds have been released to the global environment from human activities. Among these are synthetic halogenated aromatic hydrocarbons (HAH) which are extremely persistent¹, widely distributed in the environment² and, due to their high hydrophobicity, exhibit potential for bioaccumulation and biomagnification.³⁾ In recent years, increasing attention has been focused on measuring HAHs in marine coastal environments and especially in fish and shellfish harvested from these waters, in general, many of the contaminants introduced originally in the water or deposited from the atmosphere have a high affinity for the particulate phase and sediments generally constitute a major sink for lipophilic organic contaminants. However, because sediments are also a biological habitat, food web uptake of these toxicants is strongly influenced and driven by their presence in the abiotic compartment. The toxicological and environmental chemistry of polychlorinated biphenyls (PCBs), dibenzo-p- dioxins (PCDD) and dibenzofurans (PCDF) has been extensively investigated.⁴⁻⁵⁾ Among the HAHs, in addition to the PCDDs and PCDFs, are the AHH active PCB congeners. These compounds, which resemble 2,3,7,8-TCDD in their mechanism of action, have been widely recognized as responsible for many of the adverse effects observed in wildlife. Since AHH active PCBs may be found in environmental samples at levels orders of magnitude greater than PCDD/Fs, their presence can not be neglected when establishing the risk associated with 'dioxin like" contaminants. However, the analysis of the most toxic and thus more relevant PCBs has been generally limited to the determination of the planar or non-ortho congeners and some of their mono-ortho derivatives $6-7$). This investigation reports the determination of dimono- and non-ortho PCB congeners as well as the seventeen 2,3,7,8-substituted PCDD/PCDF in sediments and selected organism from Galveston Bay , Texas.

Experimental

Surticial sediment samples were collected from a small boat using a hand held box coring device at depths ranging from 2 to 3 meters. All efforts were made to sample line grained sediments from every station. After collection, all sediments were stored in clean combusted glass containers and kept

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refrigerated until arrival at the lab. All samples where then logged in and frozen until the time of analysis. American Oysters (*Crassostrea virginica*), Blue Crabs (*Callinectes sapidus*), and Hardhead Catfishes (Arius felis) were collected by manual dredging, gill net and hook and line, respectively. All organisms were dissected and the section of tissue to be analyzed was homogenized and kept frozen at -20°C until time of analysis. Collection was conducted in Galveston Bay, Texas at or near five locations regularly sampled for the NOAA Status and Trends Program : Carancaliua Reef (GBCR), an oyster reef located in East Bay, Hanna"s Reef (GBHR), a large commercial oyster reef located in West Bay, Todd's Dump (GBTD), a small oyster reef off Eagle Point, the Houston Yacht Club (GBYC), and the Houston Ship Channel (GBSC) near Morgan's point in Tabb's Bay (Figure 1). The analytical protocol used for the separation and determination of the targeted analytes was described elsewhere.⁸⁾ Briefly, biological tissues were extracted by macerating them with methylene chloride using a tissue homogenizer while sediments were extracted with toluene using a sohxlet dean stark (SDS) assembly. To eliminate interferences caused by lipid material, the hexane extract was treated with a mixture of concentrated sulfuric acid and silica gel. The sample extracts were further purified by mixed-bed silica column chromatography (8 g of 44% sulfuric acid/silica gel and 4 g of 33% 1 M sodium hydroxide/silica gel) and basic alumina column chromatography. Sample extracts were then fractionated at ambient pressure using a Michael-Miller type chromatographic column. The column was packed with 2 g of a 20:1 charcoal/silica gel adsorbent mixture and sequentially eluted with 60 ml of 1:4 methylene chloride : cyclohexane, 40 ml of 1:9 methylene chloride : toluene and 25 ml of toluene to isolate the di-, mono-, and non-ortho PCB congeners in fractions 1, 2 and 3, respectively. The fraction containing the PCDD/PCDF was then recovered by reversing the column and eluting with 200 ml toluene All four fractions collected from the charcoal column were then analyzed individually by high resolution mass spectrometry using a VG Autospec Ultima (VG Analytical. Manchester ,U.K.) double focusing instrument working at >10,000 resolving power (10% valley) operated in electron impact mode (El) at 36eV coupled to a HP-5890 SeriesII gas chromatograph (Hewlett-Packard, Wilmington, Delaware, U.S.A.) equipped with a 60 m x 0.25 mm id. DB-5MS fused silica capillary column (J&W Scientific, Folsom, California, U.S.A).

Results and Discussion

The geographical distribution of HAHs in the Galveston Bay is heavily dominated by thc intense urban and industrial activity located along Buffalo Bayou in the upper portion of the Houston Ship Channel.⁹⁾ In general, a concentration gradient is observed for most of the contaminants as sampling stations progress seaward and away from the upper portion of the bay. The concentrations and distributions of the analytes targeted in this study are presented in Table 1 and 2. Total PCDD/PCDF expressed as the sum of all the 2,3,7,8-substituted isomers was much higher in the abiotic compartments than in any of the organisms analyzed. Sediment distributions were mainly comprised ofthe heavier, highly chlorinated PCDDD/PCDF isomers (i.e. OCDD). Nevertheless, 2,3,7,8-TCDD and TCDF were detected in most sediment samples at concentrations ranging from non detected (nd.) to 4.44 pg/g and 0.650 to 7.46 pg/g respectively. In contrast, these two isomers are much more concentrated in the tissues of the analyzed organisms. 2.3,7,8-TCDD concentrations ranged from 1.15 to 16.4 pg/g in the American oysters, 1.10 to 138 pg/g in the Blue crabs and 11.1 to 101 pg/g in the Catfish livers while 2,3,7,8-TCDF concentrations were between 2.31 and 50.6 pg/g in the American oysters, from 5.77 to 385 pg/g in the Blue crabs and 5.57 to 8.04 pg/g in the Catfish livers. Although these congeners appear to be highly concentrated by thc organisms, OCDD was still detected in most of the species at concentrations ranging from 12.5 to 172 pg/g. Among species, Blue crabs contained the highest concentrations of the lower molecular weight PCDD/Fs when compared with the oysters

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and the catfish livers. The PCB distribution was dominated by the presence of congeners 138, 118, and 105 and to a minor extent 128,158,156, 170 and 77. However, in the Ship Channel organisms all targeted PCB congeners were detected including some which are only found at very low concentrations in Aroclor mixtures such as PCB 114, 123, and 167. Although extensive bioaccumulation was observed for all species, contrary to the trend noticed for the PCDD/PCDFs the concentrations of the toxic PCB congeners in the catfish livers were much more elevated than in the crabs and oysters. In order to simplify the interpretation of the data, the 2,3,7,8-TCDD toxic equivalents (TEQs) were calculated using the available toxicity equivalency factors (TEF).⁵⁾ However, for those PCB congeners where no TEF was available a more conservative approach was used and the value for a representative congener was used (i.e the TEF for PCB 77 was used for PCB 81). TEQs concentrations varied as follows: levels between 8.68 to 50.6 pg/g were observed for oyster samples, between 13.9 and 422 pg/g for the Blue crab tissues, and between 44.4 to 424 pg/g for the Catfish livers. In contrast, the TEQs observed for the sediments ranged from 2.04 to 10.12 pg/g . While there is a clear increasing trend for TEQ concentrations between organisms from the lower bay and the Ship Channel, the sediment TEQs does not show a significant trend. The individual contributions of the HAHs to the TEQ for all samples analyzed from Todd's Dump, a station located in the middle of Galveston Bay, are shown in Figure 2. The graphs indicate that the major contribution to the TEQ for the sediments is due to the presence of high molecular weight PCDD/PCDFs. However, the non-ortho PCB congeners appear to play a major role in the TEQs observed for the biological samples (29 to 74%). In addition, in spite of their relatively high concentrations, thc di-ortho PCB congeners have a very limited or negligible contribution to the observed TEQs (<2%).

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Figure I - Sampling sites along the Galveston Bay System.

Figure 2.- Percentage contribution of non-, mono, -di-ortho PCB congeners; PCDF and PCDD to die 2,3,7,8-TCDD Toxic Equivalents (TEQs) in samples from the Todds Dump site.

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Table 1.- Concentrations (pg/g) of the seventeen 2,3,7,8-substituted PCDD/PCDF in selected organisms and sediments from Galveston Bay, Texas.

nd.: non detected

(0.355): the number is below the method determination limit.

Table 2.- Concentrations (pg/g) of non-, mono- and di-ortho PCB congeners in selected organisms and sediments from Galveston Bay, Texas.