Comparative Toxicity in Developmental Stages of Fish from 2,3,7,8-Tetrachlorodibenzo-*p*-dioxin (TCDD)

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1. Introduction

A number of studies have shown that the toxic effect of TCDD depends on the fishes life stage and species. In order to properly assess the impact of TCDD in the environment it is important to know the TCDD body burden. Little is known about the relationship between the TCDD body burden and the toxic effects of TCDD in fish^{1.2.3)}.

This paper compares the sensitivity and toxicity of TCDD in early life stages of the fathead minnow (*Pimephales promelas*) with other fish species. In addition, TCDD toxicity studies were carried out using the fathead minnow embryo and larvae as an animal models which emphasize on the relationship between the tissue dose and the biological effects.

2. Methodology

Embryos and larvae of the fathead minnow (*Pimephales promelas*, Rafinesque) were waterborne exposed to [3 H]2,3,7,8-tetrachlorodibenzo-*p*-dioxin (TCDD) (40 Ci/mmol, radiochemical purity of 98%). Fish embryos and larvae were divided into a no treatment control, a solvent vehicle control (acetone <0.05%), and 5 concentrations of TCDD. The TCDD concentrations in water and in tissue were measured by a scintillation counter (Tracor Analytic Mark III) with a detection limit of 0.001 ng/g. Fathead minnow embryos were exposed at the blastula stage in a static non-renewal system. The concentrations of TCDD at the time of embryo exposure were 0.37, 0.59, 1.2, 2.83 and 10.16 ng/L. The embryos were examined daily through development for the appearance of lesions and up to 48 hs after hatching. Larvae of one month old (0.0017-0.0094) were exposed for 24 hr to TCDD at concentrations of 3.8, 40.0, 68.6, 78.48 and 90.0 ng/L. Then, larvae were transferred into clean water tanks and sacrificed at day 64 after exposure. Growth, appearance of a wasting-type syndrome, and death were noted.

3. Results and Discussion

The lowest adverse effect level (LOAEL) in the embryos of the fathead minnow was shown at a tissue dose of 0.04 ng/g and the LD50 (the measured TCDD tissue dose in eggs or embryos that caused 50% of death) was at a tissue dose of 25.71 ng/g TCDD (wet weight)⁴). When compared these findings with the ones in lake trout, the sac fry stage had 23% mortality at a TCDD tissue dose of 0.040 ng/g, an ER50 (residual concentration that produces effect on

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50 % of the population) of 0.065 ng/g, and a LOAEL of 0.055 ng/g (48 hr of exposure)¹⁰. Gross lesions such as multifocal hemorrhages, edema, and deformation of the lower jaw were observed in the fathead minnow. These lesions were classified as mild and severe due to their onset and their pathogenesis. Other toxicity studies have reported similar types of lesions in the rainbow trout, lake trout, northern pike, mummichog and medaka embryos³⁰.

Fathead minnow larvae exposed to TCDD had a non-observable adverse effect level (NOAEL) at a tissue dose of 4 ng/g. Significant differences in growth and wasting-type syndrome were shown at a tissue dose of 20 ng/g, and 100% mortality at 163 ng/g (wet weight)⁵⁾. The tissue doses are in a similar range to earlier reports on fathead minnows of 0.5-1.5 g, in which 100% mortality was found with 2.4-1,280 ng/g TCDD (dry weight)⁶⁾. Signs of toxicity such as fin necrosis, body weight loss and lethargy in fathead minnow larvae are similar to other studies in which the same or different species, systems anc/or experimental designs were applied³⁾.

The comparison of the results in the present study with other acute exposures of embryo and larvae < 8 g are summarized in Table 1. Table 1 is a modification of Table 4-1 that appeared in the EPA 1993 Interim Report³⁰. It has been updated with the latest studies on the effects of TCDD in early life stages of fathead minnows.

4. Conclusions

a) The relative toxicity to the early life stages in fish based on body burden are: embryo>larvae>juveniles. Adult fish are the least sensitive stage to TCDD.

b) In a number of fish species, the toxicity in embryos is associated with vascular lesions which may determine the death of the organism, while the toxicity in larvae is associated with a decrease in growth and the occurrence of a wasting-type syndrome.

c) The fathead minnow embryos and larvae (0.0017-0.0094 g) are one of the most sensitive fish species to TCDD based on body burden.

d) The LOAEL in fathead minnows embryos was 0.04 ng/g, and the NOAEL in larvae was 4 ng/g TCDD. These concentrations in feral fish have only been observed in highly contaminated areas. Therefore, it is unlikely that in less contaminated areas, fathead minnow body burdens would reach toxic levels.

Test Species	Test Method	Water	Organism	Duration		Effect	Kelerence ⁶
		(`onc. (ng/l_)*	('one. (ng/g)*	Ехромите	Observation		
eggs	C88		0.230	#ingle	Fertilized egg	LR50*	Walker and
McConaughy strain	inject.		(in eggn)	injection	to swim-up	:uc fry	Peterson 199
~EE*	°88		0,240	single	Festilized egg	_K50"	Walker and
Erwin strain	inject.		(in eggs)	injection	to wim-up	inc fry	Peterson 199
cgga	<88		0.374	single	Fertilized egg		Walker and
Arlee strain	(injection)		(in eggs)	injection	to swin-up	raic try	Peterson 199
cggs	cgg		0.488	single	Fertilized egg	LR504	Weiker and
Fagle Lake strain	(injection)		(in eggs)	injection	to swim-up	Hac fry	Peterson 199
cggs	~88		0.421	single	>48h to post	1.850*	Walker e: al.
Fish Lake strain	(injection)		(in eggs)	injection	swim-up	tak fry	1992

Table 1. Literature summary of the toxicological effects during the acute exposures on fish embryos and larvae to 2,3,7,8-tetrachlorodibenzo-*p*-dioxin. (Modified from EPA 1993³)

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continuation	<u> </u>						
Continuation Tent Species	Test Method	Water Conc. (ng/[.)*	Organism Conc. (ng/g) ^b	Duration Exposure	Observation	Effect	Reference*
Rainbow trout			· · · · ·				
cggs	water (renewaj)		0.279 (in eggs)	48h	>48h to post swim-up	Significant mortality in suc fry	Walker <i>et al.</i> , 1992
588ª	water (renewal)		0.439 (in c ggs)	48h	>48h to post awiin-up	LR50 ⁴ suc fry	Walker <i>et al.</i> , 1992
win-up fry (0.38 g)	water (flow-tijru)	0.176	3.22	28d	284	95% mort∎lity	Mehrle <i>et al</i> , 1988
wim-up fry (0.38 g)	water (flow-thru)	0.0011	0.021	28d	284	NOAEL*	Mehrie <i>et al.</i> , 1988
wim-up fry (0.38 g)	water (flow-thru)	0.038	0.765'	28d	284	LOAEL* 45% mortality	Mehrle <i>et al.</i> , 1988
ingerling 3-7g)	diet (0.494 ng/	g)	0.25	13wk	13wk	No toxic effect	Kleeman <i>et al.</i> , 1986a
fingerling (8g)	i.p. injection		10	single injection	2-4wk post exposure	Fin necrossis, no effect on immune suppression	Spitsbergen <i>et</i> 1986
Lake trout (<i>Salvelinus</i> :88 ⁸	namaycush) water (renewal)		0.034 (in eggs)	48h	>48h to post swim-up	NOAEL.	Walker <i>et al.</i> , 1991
cg gs	water (renewal)		0.040 (in cgg#) .	4Xh	>48h to post awim-up	23% mortality in sac fry	Spitsbergen et al., 1991
rgg* (renewal)	water (in eggs)		0.055	48h swiin-up	>48h to post mortality	LOAEL"sae fry	Walker <i>et al</i> . 1991
:ggs (renewal)	waler (in eggs)		0.065	48h swim-up	>48h to post	1.R50" sac fry	Walker et al., 1991
eggs (injection)	cggs (in cggs)		0.047 injection	single swim-up fry	Fertilized egg to	LR50" nac fry	Welker <i>et al.</i> , 1992
Fathead minnow (Pin CEEF	nephales pro water (static)	omelas) 0.37	0.04	fertilized egg to 2d post hatch		LOAEL [*] embryon with leasons	Olivieri and Coc 1995 ⁴¹
cggs	water (static)	1.2	0.16	fertilized egg to 2d post hatch		ED50" (0.14 ng/g)	Olivieri and Coc 1995 ⁴⁰
c889	water (static)	0.59- 10.16	2.46- 37.01	fertilized egg to 2d post hatch		1.1)50" (25.71 ng/g)	Olivieri and Co 1995 ⁴¹
arvac 0.0017-0.0094g)	water (static)	3.8	4	24h	ત્મન	NOAEL*	Olivieri and Cos 1994 ^w
arvac (0.0017-0.0094g)	water (static)	90	20	24h	હ્નન	Decreased growth, wasting-type syndrom	Olivieri and Co e 1994 ⁵⁴
arvae (0.0017-0.0094g)	water (static)	40	26	24h	64ત	5.9% mortality, wasting-type syndrom	Olivieri and Co e 1994 ⁵⁶
larvac (0.0017-0.0094g)	water (static)	78.48	45	24h	644	32 % mortality, decreased growth, wasting-type syndrom	Olivieriand Co 1994 ⁵ ⊯
larvac (0.0017-0.0094g)	water (static)	68.6	163	24h	324	100% mortality, decreased growth, wasting-type syndrom	Olivieri and Co 1994 ¹⁰ e

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continuation Teat	Test	Water Conc. (ng/L)*	Отдалина	Dunitsin		Effect	Reference"
Species	Method		Conc (ng/g) ^b	Exposure	Observation		
Fathead minnow		-				<u>_</u>	
juvenile	water	81	2 4-1 280	24h	404	(*) * mortality	Adums et al
(0.5-1.5g)	(renewal)						1986
Japanese medak	(Oryzias latip	r.c)					
cggs	water		0.240	fertilized egg to Ad post hatch		ER50 ^e (embryos	Wisk and
	(sinte)		(in embryos)			with lesions)	Cooper 1990b
Yellow perch (Pe	rca flavescens)						
juvenile	diet		0.143	1.3wk	13wk	No toxic effect	Kleeman <i>et ul</i>
	0.494ng/g						19866

Measured TCDD concentration in water

^b Measured TCDD concentration in organism based on wet weight For literature citation referred to EPA 1993 Interim Report

* LR50 (corrected for control mortality) The measured residue concentration in eggs that caused 50% mortality to use try

ER50 The measured reaidue concentration in eggs that caused 50% effect

1D50 The measured TCDD tissue dose in eggs or larvae that caused 50% death. ED50 The measured TCDD tissue dose in eggs or larvae that caused 50% effect

NOAEL= No Observed Adverse Effect Level

LOAEL+ Lawest Observed Adverse Effect Level NOAEL and LOAEL values calculated for the Interim Report³

* Measured TCDD concentration in organism based on dry weight

5. References

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